

Lifetime Reproductive Success of Female Mountain Gorillas

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ABSTRACT Studies of lifetime reproductive success (LRS) are important for understanding population dynamics and life history strategies, yet relatively little information is available for long-lived species. This study provides a preliminary assessment of LRS among female mountain gorillas in the Virunga volcanoes region. Adult females produced an average of 3.6 ± 2.1 surviving offspring during their lifetime, which indicates a growing population that contrasts with most other great apes. The standardized variance in LRS (variance/mean² = 0.34) was lower than many other mammals and birds.

When we excluded the most apparent source of environmental variability (poaching), the average LRS increased to 4.3 ± 1.8 and the standardized variance dropped in half. Adult lifespan was a greater source of variance in LRS than fertility or offspring survival. Females with higher LRS had significantly longer adult lifespans and higher dominance ranks. Results for LRS were similar to another standard fitness measurement, the individually estimated finite rate of increase (λ_{ind}), but λ_{ind} showed diminishing benefits for greater longevity. *Am J Phys Anthropol* 000:000–000, 2011. © 2011 Wiley-Liss, Inc.

Assessments of individual fitness are important for understanding population dynamics and evolutionary biology (Stearns, 1992; Metcalf and Pavard, 2007). Fitness data for individuals can be combined to provide estimates of population growth, population viability, and the effective population size (Soule, 1987; Caswell, 2001). Variance in fitness among individuals leads to natural selection when it is linked to differences in phenotypes that are heritable (Darwin, 1859; Price, 1996; Avise and Ayala, 2009). The standardized variance (variance/mean²) in fitness has been defined as the “opportunity for selection,” which reflects an upper limit on the rate of evolution (Crow, 1958; Shuster and Wade, 2003). Variance in fitness can also arise from environmental fluctuations or by chance (Sutherland, 1985, 1987; Coltman et al., 1999; Forcada et al., 2008).

Lifetime reproductive success (LRS) is a commonly used measure of individual fitness (Barrowclough and Rockwell, 1993; van Noordwijk and van Schaik, 1999; Rhine et al., 2000; Schiegg et al., 2004). The LRS of a female equals the total number of surviving offspring that she produces in her lifetime. LRS can be expressed as the product of three components: adult lifespan (L), average fecundity per year (F), and offspring survival (S). All three components are expected to have positive correlations with their product (LRS) unless they have negative covariance amongst each other. For example, a negative correlation between fertility and adult lifespan could arise from life history trade-offs between reproduction versus somatic maintenance (Kirkwood, 1977; Kirkwood and Holliday, 1979; Hurt et al., 2006; Ricklefs and Cadena, 2007), and between genes that enhance survival or reproduction early in life at the expense of harmful effects later in life (Williams, 1957; Blomquist, 2009b). Conversely, positive covariance among all three components could arise if females in favorable environments have better physical condition that enhances many aspects of fitness (“silver spoon effect,” Grafen, 1988).

Body mass and social ranking have been considered indicators of female quality that could influence reproductive success (Ribble, 1992; Berube et al., 1999; White, 2005; Jones et al., 2010). Thus, a combination of factors can determine how each component contributes to variance in fitness (Brown, 1988).

Another commonly used measure of individual fitness is the individually estimated finite rate of increase, also known as the per-projection-interval multiplicative rate of increase (McGraw and Caswell, 1996; Kaar and Jokela, 1998; Weladji et al., 2006). The individually estimated finite rate of increase (λ_{ind}) is analogous to population-wide values of λ from Leslie matrix calculations, whose natural log is equal to the population growth rate “ r ,” also known as the instantaneous rate of increase (Caswell, 2001). The value of λ is greater than one when a population is growing and less than one when a population is declining. The population growth rate would equal the natural log of a female’s value for λ_{ind} if all other adult females gave birth to the same number of surviving offspring at the same ages.

Unlike LRS, values for λ_{ind} depend upon the ages at which a female reproduces because of the discounting inherent in age-structured population models (Caswell, 2001). Among females with the same LRS, earlier reproduction increases λ_{ind} when it is above one, and delayed reproduction increases λ_{ind} when it is below one. Thus, early breeding is generally considered favorable (i.e., it

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leads to a greater number of descendants over multiple generations) when a population is increasing, and delayed breeding is considered favorable when a population is declining (Stearns, 1992; McGraw and Caswell, 1996; Brommer et al., 2002; but see Dugdale et al., 2010). When λ_{ind} is above one, the driving force for longer adult lifespans can become minimal, allowing deleterious mutations to accumulate, and senescence can evolve through “decelerating selection” (Medawar, 1946; Hamilton, 1966; Laird and Sherratt, 2009).

Species with longer adult lifespans typically have lower fertility, a combination known as “slow” life history (e.g., Charnov and Berrigan, 1993; Leigh and Blomquist, 2007). The life histories of great apes are among the slowest of all primates, a correlation which is consistent with their large body sizes (Purvis et al., 2003; Deaner et al., 2003). Studies of life history theory often involve inter-specific comparisons of demographic parameters, so data on great apes are relevant for understanding primate life history evolution generally and that of humans in particular (Leigh and Shea, 1996; Dunbar, 2003; Hawkes, 2003).

Despite the importance of LRS and λ_{ind} for studies of evolutionary biology and population dynamics, relatively few direct measurements are available for long-lived species, especially nonhuman primates (Rhine et al., 2000; Wasser et al., 2004). The lack of data reflects the challenges of tracking individuals in long-term studies, especially when those individuals disperse (e.g., Alberts and Altmann, 1995). In the absence of data that spans the entire lifetime of all individuals in a cohort, LRS and its variance are often estimated from measurements of short-term reproductive success (Rhine, 1992; Barrowclough and Rockwell, 1993; Storz et al., 2002; Cheney et al., 2004; Setchell et al., 2005; Robbins, 2010).

This paper provides a preliminary assessment of LRS and its variance among female mountain gorillas in the Virunga volcanoes region, including comparisons with λ_{ind} , partitioning into components (L , F , and S), and the use of dominance rank as a potential indicator of individual quality. The population of those groups has grown by 3–5% per year throughout this study, so we expect that λ_{ind} will typically be above one, and it will place diminishing value on reproduction later in life (Miller et al., 1998; Robbins and Robbins, 2004; Robbins, 2010). However adult females are almost always nursing or pregnant, so we see relatively little variation in how the timing of their reproduction is distributed throughout adulthood (Robbins et al., 2006; Robbins et al., 2011), and therefore we expect a close correlation between LRS and λ_{ind} .

Female mountain gorillas feed on abundant and evenly distributed vegetation, so variability in food density (from 4 to 66 g/m²) and group size (from 2 to 65 individuals) have shown no significant influence on female reproductive success (Robbins et al., 2011). There was no evidence of leopard predation on gorillas during this study and we are unaware of any sightings of those predators since the 1960s (Schaller, 1963). Human disturbance, while considerable, has not substantially influenced fertility or offspring survival (Robbins et al., 2011), and our analysis controls for the effects of poaching. Mountain gorillas and other primates have relatively low variance in their annual rates of fertility and mortality, which may reflect low environmental stochasticity and/or enhanced buffering against such stochasticity (Morris et al., 2011). Therefore, we predict that female mountain gorillas have low standardized variance

TABLE 1. Summary of the study groups

Group	Year 1	Year 2	Grp-years	Size	AF	Births
Amok's Grp	1969	1971	2.3	1.1	0.1	0
Beetsme's Grp	1985	2008	23.5	19.3	6.1	33
Bwenge	2006	2008	2.1	8.4	5.2	7
Group 4	1967	1979	11.3	11.0	4.0	12
Group 5	1967	1993	25.7	17.6	6.6	44
Group 8	1967	1974	6.7	4.2	0.7	1
Gwiza	2004	2008	4.8	1.1	0.1	0
Inshuti	2004	2008	5.1	1.7	0.6	1
Isabukuru	2007	2008	1.6	6.2	3.0	2
Kuryama	2007	2008	1.5	12.6	4.9	3
Nunkie's Grp	1972	1985	12.6	9.5	4.3	15
Pablo's Grp	1993	2008	15.7	42.2	15.5	67
PbSubgrp	2007	2007	0.1	14.0	3.0	0
Samson's Grp	1971	1976	4.6	1.6	0.6	1
Shinda's Grp	1993	2008	15.7	22.3	6.6	26
Susa Grp	1978	2008	30.4	27.7	9.5	75
Tiger's Grp	1981	1987	6.1	1.5	0.3	1
Umushikirano	2006	2008	2.3	1.2	0.2	0

First and last years of observation (year 1 and year 2), as well as the number of gorilla-years observed (grp-years). Average number of total gorillas (size) and adult females (AF) throughout the study. Number of births in each group.

(variance/mean²) in LRS, especially when poaching deaths are excluded from the calculations.

Female mountain gorillas form weak individualistic dominance hierarchies and competition for resources is generally ineffective (Stewart and Harcourt, 1987; Watts, 2001; Robbins et al., 2005). Female mountain gorillas had shorter interbirth intervals when high ranking, but rank was not significantly correlated with offspring survival or the annual rate of giving birth to surviving offspring (i.e., the product of fecundity times offspring survival, Robbins et al., 2007a,b). Those mixed results suggest that dominance rank may represent an approximate indication of differences in physical condition among females. This study builds upon those results by examining the relationships between rank versus adult lifespan, LRS, and λ_{ind} ; and by looking for positive covariance among all reproductive parameters as additional evidence of differences in female quality.

Studies of individual fitness should ideally last several times longer than the lifespan of typical adults, which would take hundreds of years for long-lived species such as mountain gorillas. Even though this 40-year study is one of the larger ones for nonhuman primates, sample sizes are still small, so conclusions from this preliminary assessment should be considered tentative. It is also important to consider how the modern environment of mountain gorillas has been influenced by human activity (including the near-extirpation of nonhuman predators), although a full discussion of such influences is beyond the scope of this publication.

METHODS

Study population

The study population consists of mountain gorillas in the Virunga Volcano region of Rwanda, Uganda, and Democratic Republic of Congo. The data are taken from the long-term records of the Dian Fossey Gorilla Fund International (DFGFI) Karisoke Research Center from 1967 to 2008, along with one group (Susa) from the records of the Rwanda Development Board (Table 1). More groups have been followed per day later in the study than earlier. Female gorillas are considered

TABLE 2. Summary of lifetime reproductive success (LRS) and λ_{ind} for each female, and the components of LRS: adult lifespan in years (L), average fecundity per year (F), offspring survival (S), and their products (LF , LS , and FS)

Gorilla	IBIS	F	S	L	LF	LS	FS	LRS	λ_{ind}
CLO	43.4	0.308	71%	22.8	7.0	16.3	0.220	5.0	1.057
FUD	56.8	0.226	83%	26.6	6.0	22.2	0.188	5.0	1.049
GIT	—	0.273	33%	11.0	3.0	3.7	0.091	1.0	0.950
IMP	—	0.364	0%	2.8	1.0	0.0	0.000	0.0	—
JEN	49.6	0.225	100%	13.3	3.0	13.3	0.225	3.0	1.032
MAC	—	0.226	50%	8.8	2.0	4.4	0.113	1.0	0.950
MUR	46.0	0.239	75%	16.8	4.0	12.6	0.179	3.0	1.025
PAN	47.8	0.224	83%	26.8	6.0	22.4	0.186	5.0	1.043
PTY	50.8	0.282	71%	24.8	7.0	17.7	0.201	5.0	1.051
PAP	43.9	0.253	78%	35.6	9.0	27.7	0.197	7.0	1.054
PIC	47.6	0.179	100%	16.8	3.0	16.8	0.179	3.0	1.027
PUC	51.1	0.231	86%	30.3	7.0	26.0	0.198	6.0	1.054
SIM	50.2	0.382	33%	23.6	9.0	7.9	0.127	3.0	1.022
Main dataset									
Mean	48.7	0.262	67%	20.0	5.2	14.7	0.162	3.6	1.026
Std	3.9	0.059	29%	9.4	2.6	8.8	0.063	2.1	0.038
G'	—	14%	53%	60%	70%	120%	36%	—	—
R^2	1%	8%	44%	93%	68%	92%	63%	—	79%
No poaching									
Mean	49.0	0.258	74%	23.2	6.0	17.4	0.181	4.3	1.034
Std	4.1	0.057	24%	7.6	2.3	7.6	0.041	1.8	0.032
G'	—	30%	64%	67%	93%	124%	30%	—	—
R^2	4%	2%	18%	86%	46%	86%	43%	—	74%

The age of death is 8 years more than the adult lifespan. IBIS is the average length of inter-birth intervals (months) with offspring that survive to reach age 3, LF equals the total number of offspring per female, FS equals annual rate of giving birth to offspring that survive to reach age 3, and LRS equals the number of offspring that survived to reach age 3. The G' and R^2 values indicate how each component contributes to variance in LRS (see Methods).

infants until they reach the age of 3 years, then juveniles until age 6, then subadults until age 8, and then adults thereafter.

The “main dataset” in this study consists of 13 breeding females that were monitored for essentially their entire adult lifespan (Table 2). Eleven of those females definitely or probably died, and results were similar whether we included or excluded the other two females (Clo and Jen) whose disappearances were unexplained (not shown). We always excluded two additional adult females who disappeared for unknown reasons while nulliparous. These disappearances were most likely due to dispersal, given that 31 known or probable transfers by nulliparous adult or subadult females occurred through 2005 (including transfers among the study groups, Robbins et al., 2009a) and only one known death. One female (Pan) was estimated to be 10 years old when first observed, so she could have already had an infant that died before weaning, but results were similar when we added another deceased offspring to her records (not shown).

Lifetime reproductive success

The LRS of each female in the main dataset ($N = 13$) was defined as her total number of offspring that survived to reach the age of 3 years, the typical age of weaning in this population (Watts, 1991). Values of LRS do not include any offspring that died during infancy (i.e., before age 3), regardless of their cause of death. LRS values include all offspring that survived through infancy, regardless of their subsequent fate as juveniles and beyond.

We report the mean, standard deviation, and standardized variance (variance/mean²) in LRS for the main dataset, which includes three adult females that died from poaching (Imp, Mac, Mur). It is unclear whether

poaching deaths are typical of the evolutionary history of this population (Harcourt and Stewart, 2007), so we performed a separate set of LRS calculations without those females. The remaining 10 females did not have any offspring that died from poaching during infancy. We performed an ANOVA to quantify how the poaching deaths of adult females contributed to variance in LRS in the main dataset. In that analysis, the dependent variable was the LRS of each female in the main dataset. The category variable was whether the female died from poaching (“yes” or “no”). The coefficient of interclass correlation (r_i or R^2) from that ANOVA indicated how much variance in the dependent variable (LRS) could be explained by the category variable (e.g., Sokal and Rohlf, 1995).

Other measures of lifetime fitness

Studies of LRS often consider offspring survival until adulthood (Brown, 1988; Strassmann and Gillespie, 2003), so for comparison purposes, we estimated the average value for the females in the main dataset. We multiplied the average LRS of those females by the overall survivorship from ages 3 to 8 of all habituated gorillas in the Virungas (Robbins et al., 2011). We performed a similar calculation using the LRS of the 10 adult females who did not die from poaching, and a survivorship value for ages 3–8 that excluded poaching deaths at those ages (Robbins et al., 2011).

Studies of female LRS occasionally focus on the number of daughters that survive to reach adulthood, which reflects the capacity for one generation of females to replace itself with another generation of females (Kaar and Jokela, 1998; Brommer et al., 2004). The birth sex ratio of this population has not been significantly different from 50:50 (Robbins et al., 2007a), and the probabil-

ity of dying at each age until adulthood has not been significantly different for male versus females (Robbins et al., 2011). Therefore, to avoid any bias from stochasticity in our relatively small dataset, we assumed that the average number of “replacements” per adult female was equal to 50% of the average number of total offspring that survived to reach adulthood.

We calculated the individually estimated finite rate of increase (λ_{ind}) for each female, using an individual-based discrete-time renewal equation:

$$1 = 0.5 * \sum_{x=1}^{\Omega} f_x * s_x * (\lambda_{\text{ind}}^{\wedge -x}) \quad (1)$$

in which “ f_x ” is the number of offspring that the female produced at each age “ x ” until the last observed age “ Ω ,” and s_x is the proportion of those offspring that survived to reach the age of 3 years. Calculations of λ_{ind} often do not account for variability among females in the sex ratio of their offspring (McGraw and Caswell, 1996; Brommer et al., 2004), so the coefficient of “0.5” reflects the assumption of a 50:50 male: female ratio throughout the population (Robbins et al., 2007a). For each female, we used iterative calculations with the bisection method to find the value of λ_{ind} that corresponds to a value of “one” for the right-hand side of Eq. (1) (Byrne, 2008).

We performed an ANOVA to quantify the proportion of variance in λ_{ind} among females that could be explained by differences in their LRS, versus the proportion that arose from differences in the timing of reproduction among females that had the same LRS. In that analysis, the dependent variable was the λ_{ind} of each female and the category variable was her LRS. The analysis used the main dataset that includes females who died from poaching, because we are mainly comparing mathematical relationships that are not expected to depend upon the cause of death.

Our results for λ_{ind} do not include the one adult female (Imp) that was killed before weaning any offspring, because when $\lambda_{\text{ind}} = 0$, it was many (~ 25) standard deviations away from the mean of the remaining data (for a similar approach see Kaar and Jokela, 1998). If that data point was included, it would further increase the proportion of variance in λ_{ind} that arose from differences in LRS, and further decrease the linearity of the relationship between adult lifespan and λ_{ind} .

Contributions to variance in LRS

The LRS of each female equals the product of three components: her adult lifespan (L), her average fecundity per year (F), and the proportion of her offspring that survived to reach the age of 3 years (S). Adult lifespan equals the age of death minus the age when females become adults (8 years). The average fecundity of each female equals her total number of offspring (all offspring, not just those that survived to reach the age of 3 years) divided by her adult lifespan. Offspring survival for each female equals her number of offspring that survived to reach the age of 3 years, divided by her total number of offspring (all offspring). Female mountain gorillas have shorter interbirth intervals after an infant dies (Gerald, 1995; Robbins et al., 2006), so we use the average length of interbirth intervals with surviving offspring (IBIS) as an additional measure of female fertil-

ity, as well as the rate of giving birth to offspring that survive to reach the age of 3 years ($F \cdot S$).

To quantify how L , F , and S (and their products) contribute to variance in LRS among breeders, we report the G' values from upward partitioning as presented in Table 27.1 of Brown (1988). G' is an indicator of how the variance in a parameter contributes to the variance in its product with other parameters (cf. references in Breuer et al., 2010, for additional literature about this approach and others). The first step in our calculations was to normalize the data for each component, by dividing the value for each female by the mean of the component among all females. For example, if Y_L is the array of values for adult lifespan (one value for each female), and X_L is the array of normalized values, then $X_L = Y_L / \text{mean}(Y_L)$. Next we calculated the product of all three normalized components for each female (i.e., $X_L \cdot X_F \cdot X_S$, in which X_S and X_S are the arrays of normalized values for fecundity and offspring survival). Then, we calculated the variance for each of the normalized components and their product. For example,

$$G_L = \text{Var}(X_L) \quad (2)$$

$$G_{\text{LFS}} = \text{Var}(X_L \cdot X_F \cdot X_S) \quad (3)$$

in which “Var” signifies the variance among females for the value in parentheses. G_F and G_S were calculated with an analogous form of Eq. (2); G_{LF} , G_{LS} , and G_{FS} were calculated with an analogous form of Eq. (3). Finally, we multiplied each of the G values by $(100/G_{\text{LFS}})$ to produce the G' values.

In addition to the G' values, we also used linear regressions to quantify how each component and their products contributed to the variance in LRS among breeders. In each regression, the dependent variable was the LRS of each female, and the independent variable was one of the components of LRS (L , F , S) or their products. The coefficient of determination (R^2) from those regressions indicated how much variance in the dependent variable (LRS) could be explained by the independent variable (e.g., Sokal and Rohlf, 1995). For example, when the LRS of each female was the dependent variable and the adult lifespan of each female was the independent variable, the R^2 value from a linear regression equaled 0.93, which indicated that adult lifespan could explain 93% of the variance in LRS (as shown for the “main dataset” in Table 2).

Neither the G' nor R^2 values for the three components will typically add up to 100%, due to covariance among the components. The G' value for each component shows how the variance in LRS would change, if that component varied over its observed range while the other components were held constant at their average values. The R^2 values show how variance in each component affects LRS while the other components also vary over their observed range. Thus, any effect of covariance between two components is not included in the G' value for either component, but it is included in the R^2 values for both components.

We examined covariance among the components when needed to interpret their G' and R^2 values. Although we used linear regressions to describe variance contributions, we used nonparametric Spearman rank correlations for testing the statistical significance of the relationships among variables (due to small sample sizes). P -values for the Spearman rank correlations were deter-

mined by using bootstrap calculations with 10,000 permutations (Mundry and Fischer, 1998).

No infertile females have been observed in this population, so we did not use the Brown (1988) equation for estimating the variance in LRS that arises between breeders versus nonbreeders. Only one “nonbreeding adult” has been observed, a female that died from disease 3 months after reaching the age of 8 years (Qin, see Fossey, 1983). Thus her contribution to variance in LRS essentially arises from her very short adult lifespan, which is considered in the “broader” and “broadest” datasets that are described next.

Bias estimation for adult lifespan

This study did not follow a prespecified cohort of females through their adulthood, so we expected results to be biased toward females that died relatively young (e.g., Nishida et al., 2003). To estimate the bias, we compared the survivorship curve derived from the main dataset ($N = 13$) with more complete survivorship curves that included censoring of females that were still alive when the study ended. A “broader” dataset added 63 other adult females that were observed since reaching the age of 8 years but did not die during this study. The “broadest” dataset added another 20 females whose adult lifespan had already begun when first observed, including some that died during the study. Age estimates are considered accurate to within ± 2 years for females in the broader dataset and up to 4–10 years for females in the broadest dataset. The comparisons excluded the unexplained disappearances of nulliparous females because they most likely dispersed, as well as poaching deaths, so those females were also censored at the age when they were last observed.

For each survivorship curve, the period mortality probability (q_x) at each age (x) was calculated as the number of deaths (d_x) divided by the number of adult females that reached that age (Stearns, 1992). The period survival probability (p_x) equaled $1 - q_x$. Survivorship (l_x) to reach age x was calculated as the product of p_x from all preceding ages. The mean and variance in adult lifespan was estimated for each survivorship curve by assuming that the proportion of adult females that die at each age (T_x) is equal to $l_x - l_{x+1}$. The mean total lifespan (M_{TL}) was calculated as $(\sum T_x * x)$, which was converted into mean adult lifespan (M_{AL}) by subtracting the age when females are first considered adults (8). The variance in both measures of lifespan (V_L) was calculated as $\sum T_x * (M_{TL} - x)^2$.

Dominance rank

Dominance data came from fifteen hierarchies measured from 1971–2001 involving all of the major groups except the tourist group (Robbins et al., 2005). Ordinal rankings were developed from a combination of focal and ad libitum observations of approach-retreat interactions (e.g. Watts, 1994) using the I & SI method (de Vries, 1998). We standardized the rankings using Eq. (4):

$$\text{Standardized rank} = (N_F - R_F) / (N_F - 1) \quad (4)$$

In that equation, R_F is the ordinal rank to be standardized, and N_F is the number of females in the matrix. The standardized rank of each female equals the proportion of other females that are below her in the matrix, so

the lowest female in each matrix has a standardized rank of “0,” and most dominant female has a standardized rank of “1” (for similar approaches see Rhine et al., 1989; Cote, 2000). The overall standardized rank for each female was calculated as the average of all measured values throughout her adulthood. On average, the overall standardized ranks are based on 50% of the years that a female was an adult (“% years ranked,” range = 25–93%). The overall standardized ranks may be less reliable when based upon a relatively small proportion of a female’s adulthood, but when we used the “% years ranked” as a weighting factor to make the analyses less sensitive to those females, results were similar and are not shown (for a description of weighting factors, see Chatterjee and Price, 1991).

RESULTS

Lifetime reproductive success

Among the 13 breeding female mountain gorillas that were observed for essentially their entire adulthood (the “main dataset”), LRS averaged 3.6 ± 2.1 SD offspring that survived to reach age three (range = 0–7, median = 3). The standardized variance equaled 0.34. When we excluded the three females that died from poaching, LRS averaged 4.3 ± 1.8 surviving offspring, which equals a standardized variance of 0.17. Thus poaching decreased the mean LRS by 19% and doubled the standardized variance. According to the R^2 value from an ANOVA, poaching accounted for 38% of the variance in LRS among breeders ($F_{11,1} = 6.8$, $P = 0.024$).

Other measures of lifetime fitness

When poaching deaths that occurred from ages 3 to 8 are included, survivorship from ages 3 to 8 has been 91% among all habituated gorillas in the Virungas (Robbins et al., 2011). Therefore, for the thirteen females in the main dataset, the average LRS of 3.6 corresponds to an average of $3.6 \times 0.91 = 3.3$ offspring that survive to reach adulthood. If 50% of surviving offspring are daughters, then the average adult female has been producing $3.3 \times 0.5 = 1.64$ adult female offspring (“replacements”). When poaching deaths that occurred from ages 3 to 8 are excluded, survivorship from ages 3 to 8 has been 93% (Robbins et al., 2011). Therefore, if poaching did not occur, then the average adult female would produce $4.3 \times 0.93 = 4.0$ offspring that survive to reach adulthood and 2.0 replacements.

Among the 12 adult females in the main dataset with at least one surviving offspring, the average value for λ_{ind} was 1.026 ± 0.038 (range = 0.95 – 1.057, median = 1.037). The R^2 values from an ANOVA indicated that differences in LRS could account for up to 99% of the variance in λ_{ind} among females (Fig. 1), with only 1% of the variance arising from differences in the timing of reproduction among females that had the same LRS (e.g., while including all thirteen females in the main dataset, $F_{7,4} = 174.1$, $P < 0.001$).

Components of LRS

Adult lifespan was the greatest source of variance in LRS in the main dataset and when we excluded the three females that died from poaching (Table 2). Females with longer adult lifespans had significantly higher values of LRS (Tests 3.1a and 3.1b in Table 3). The positive

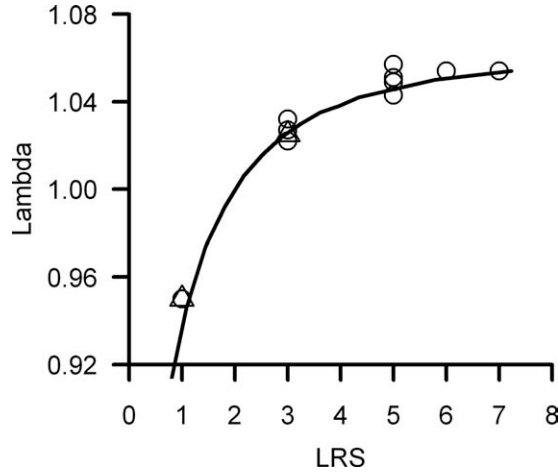


Fig. 1. LRS versus λ_{ind} (Lambda) for females with at least one surviving offspring. Not shown is one female (Imp) with no surviving offspring (LRS = λ_{ind} = 0), who was killed shortly after reaching adulthood.

TABLE 3. Spearman correlations between lifetime reproductive success (LRS) and λ_{ind} versus adult lifespan (L), offspring survival (S), fecundity (F), the average dominance rank (rank), and the average length of interbirth intervals with surviving infants (IBIS)

Test	Var1	Var2	Include poaching	N	rho	P
3.1a	L	LRS	Yes	13	0.95	0.0001
3.1b	L	LRS	No	10	0.91	0.0003
3.1c	L	λ_{ind}	Yes	12	0.73	0.0079
3.1d	L	λ_{ind}	No	10	0.61	0.0695
3.2a	S	F	No	10	-0.87	0.0020
3.2b	S	F	Yes	13	-0.82	0.0012
3.2c	S	L	Yes	13	0.44	0.1270
3.3a	Rank	L	No	9	0.83	0.0081 (0.0718)
3.3b	Rank	LRS	No	9	0.86	0.0045 (0.0390)
3.3c	Rank	λ_{ind}	No	9	0.85	0.0053 (0.0270)
3.4a	L	F*S	No	10	0.22	0.5416
3.4b	L	IBIS	No	9	0.20	0.6246

Statistical results include the sample size (N), the correlation coefficient (rho), and the P-value. Values in parentheses are the P-values when dominance rank was based on measurements before age 20.

correlation between λ_{ind} versus adult lifespan was significant when we included females that died from poaching (Test 3.1c in Table 3), but not when those females were excluded (Test 3.1d in Table 3). Visual inspection showed that λ_{ind} started to level off at longer adult lifespans (Fig. 2), as expected when λ_{ind} is greater than one.

We compared the main dataset with two other survivorship curves to assess the potential bias toward females that died at relatively young ages (Fig. 3). Adult lifespan averaged 23.1 ± 7.2 years in the main dataset, 25.9 ± 6.7 years in the “broader” dataset of females that were observed since they reached the age of 8 years (including censored data for those that have not died), and 25.2 ± 7.9 years in the “broadest” dataset that also includes females that had already reached adulthood when first observed. Thus the main dataset may underestimate the mean adult lifespan (and LRS) by about 10% and the precision of its variance is approximately $\pm 15\%$.

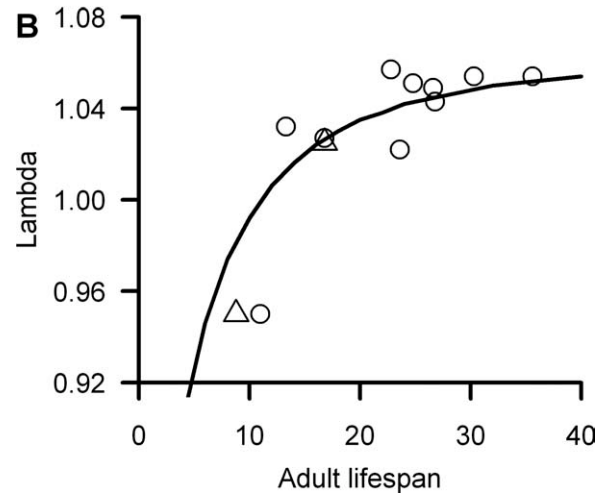
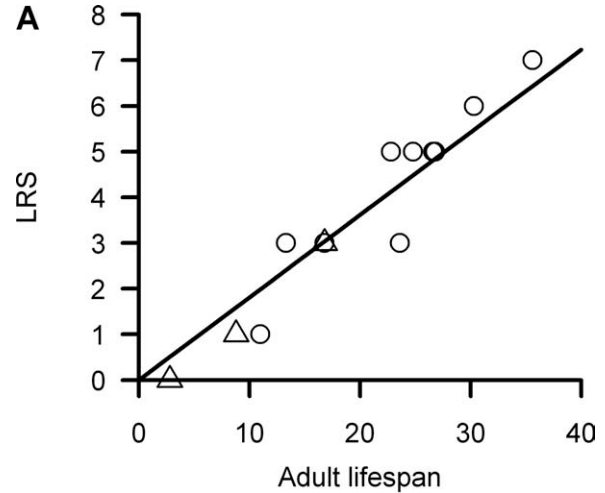


Fig. 2. Correlations between adult lifespan versus lifetime reproductive success (LRS) and the per-projection-interval multiplicative rate of increase (λ_{ind}) for females who died from natural causes (circles) and from poaching (triangles). Figure 2b does not show one female (Imp) with no surviving offspring (LRS = λ_{ind} = 0), who was killed shortly after reaching adulthood. Solid lines show hypothetical results in which the product of the other two components of LRS was held constant at the average rate of 0.162 surviving offspring per year ($F \cdot S$).

When we excluded the females that died from poaching, the G' value for offspring survival was 64% versus only 18% for the R^2 value (Table 2). This difference arose because R^2 values include the effects of covariance among components and G' values do not. Offspring survival had a strong negative correlation with fecundity (Test 3.2a in Table 3), as expected because females give birth sooner after their infants die. Thus the G' value illustrates how offspring survival would affect variance in LRS if females did not give birth sooner after an infant dies, and the R^2 value is more representative of the actual biological relationships.

When we included females that died from poaching, offspring survival still had a strong negative correlation with fecundity (Test 3.2b in Table 3), but its R^2 value was boosted by a nonsignificant positive correlation with adult lifespan (Test 3.2c in Table 3). Because of those offsetting covariances, the G' and R^2 values for offspring survival were both relatively high when we included

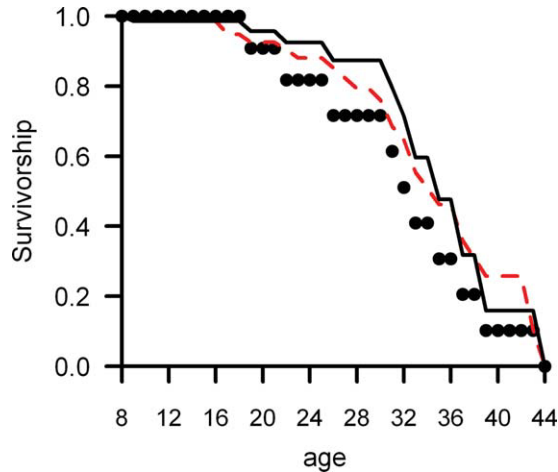


Fig. 3. Survivorship curves for adult females in the analyses of LRS (circles), versus a “broader” dataset of adult females who were observed before their first parturition (including those who are still alive, solid line), and versus the “broadest” dataset which also includes adult females who were already reproducing when first observed (dashed line). [Color figure can be viewed in the online issue, which is available at wileyonlinelibrary.com.]

poaching deaths (53 and 44%). The nonsignificant positive correlation between S and L arose because two young females (Imp and Mac) killed by poachers had infants that subsequently died and were large proportions of their total reproductive outputs. In contrast, there have been relatively few cases of offspring mortality when adult females die from natural causes (Robbins et al., 2006), and such offspring mortality would have a smaller impact on the overall offspring survival of older mothers. Among the 10 adult females in the main dataset that died of natural causes, only one lost an infant as a result of her death (Clo). Another mother had already lost her last offspring two months before she died (Sim), and the last offspring of the other eight adult females had already reached the age of 3 years when their mothers died of natural causes, and they continued to survive afterwards.

When we excluded females that died from poaching, only 1% of the variance in LRS was explained by the average length of interbirth intervals with surviving offspring (R^2 value for IBIS in Table 2). Thus fertility was the smallest source of variance in LRS. Adult lifespans were not significantly correlated with IBIS or the rate of giving birth to offspring that survive to reach age three (Tests 3.4a and 3.4b in Table 3).

Dominance rank

When we excluded females that died from poaching, females with higher dominance ranks had significantly longer adult lifespans as well as higher values of both λ_{ind} and LRS (Test 3.3 in Table 3, Fig. 4). Dominance rank typically increases with age in this population (Robbins et al., 2005), so hypothetically our results could mean that females with longer lifespans (and higher fitness) simply had more time to achieve higher ranks. When we limited the analyses to dominance data before each female reached the age of 20 years, which enabled us to compare all females over essentially the same age range, the P -values increased and the correlation with adult lifespan was no longer signif-

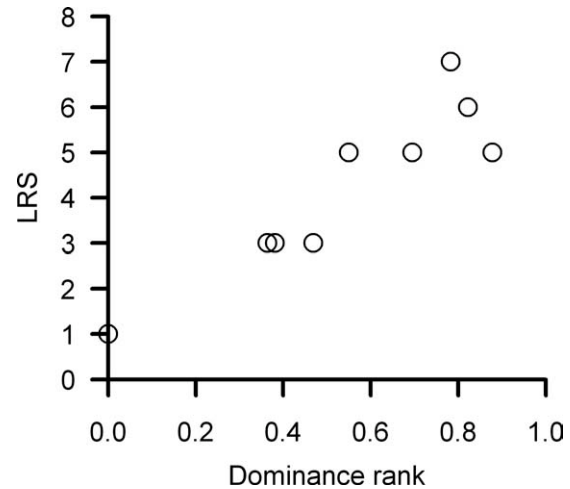


Fig. 4. LRS versus the average dominance ranking of each female. A dominance ranking of 1.0 would represent a female who had the top rank throughout her lifetime, and value of 0.0 would represent a female who had the lowest ranking throughout her lifetime.

icant (values in parentheses for Test 3.3 in Table 3). Thus, although we consistently found significant evidence that female dominance rankings may reflect differences in physical condition that affect their LRS, the potential relationship with adult lifespan is still tentative.

DISCUSSION

Lifetime fitness and its variance

This study has provided the first assessment of variance in LRS among female mountain gorillas, and one of the few such assessments for any nonhuman primate (but see Fedigan et al., 1986; Wasser et al., 2004, and references in Table 4). Despite poaching deaths, adult females produced an average of 1.6 adult female offspring during their lifetime, which indicates a growing population that contrasts with most other great apes (e.g., Sugiyama, 1994; Hill et al., 2001). Despite such recent growth within the study groups, however, the Virungas still contain only a few hundred mountain gorillas (Gray et al., 2009). The nearby Bwindi Impenetrable National Park contains a similar number of mountain gorillas at a similar density, but the two populations collectively remain a critically endangered subspecies (IUCN, 2008; Guschanski et al., 2009).

Adult female mountain gorillas produced an average of 3.6 ± 2.1 SD offspring that survived to reach the age of 3 years, which represents a standardized variance in LRS of 0.34. When we removed a major source of environmental variability from this study (poaching), the standardized variance in LRS dropped in half to 0.17. Comparable values are not available for most other populations of great apes, but the standardized variances from this study are lower than many other mammals and birds (Table 4). Thus this study illustrates that the standardized variance in LRS can become quite low in a population where major sources of environmental variability are naturally missing (no severe climate), or have disappeared (no more predators), or can be excluded from the calculations (our results without poaching). The

TABLE 4. Standardized variance (variance/mean²) in lifetime reproductive success (LRS) for female mountain gorillas versus females of other species

Common name	Genus/species	Reference	N	LRS	OFS
Mountain gorilla	<i>Gorilla beringei beringei</i>	(This study, without poaching)	10	4.3	0.17
Ache	<i>Homo sapiens</i>	(Hill and Hurtado, 1996)	n.r.	4.2	0.20
Mountain gorilla	<i>Gorilla beringei beringei</i>	(This study, with poaching)	13	3.6	0.34
California mouse	<i>Peromyscus californicus</i>	(Ribble, 1992)	16	4.7	0.35
Collared flycatcher	<i>Ficedula albicollis</i> ,	(Merila and Sheldon, 2000)	719	2.2	0.40
African lion	<i>Panthera leo</i>	(Packer et al., 1988)	47	3.8	0.41
Red deer	<i>Cervus elaphus</i>	(Kruuk et al., 2000)	202	5.2	0.41
Chacma baboon	<i>Papio ursinus griseipes</i>	(Cheney et al., 2004)	42	2.8	0.46
Rhesus macaque	<i>Macaca mulatta</i>	(Blomquist, 2009b)	208	5.9	0.47
Tawny owl	<i>Strix aluco</i>	(Millon et al., 2010)	64	6.4	0.50
Roe deer	<i>Capreolus capreolus</i>	(Vanpe et al., 2008)	28	n.r.	0.53
Bengal tiger	<i>Panthera tigris tigris</i>	(Smith and McDougal, 1991)	13	4.5	0.56
Eastern Kingbird	<i>Tyrannus tyrannus</i>	(Murphy, 2007)	159	4.4	0.65
Vervet monkey	<i>Cercopithecus aethiops</i>	(Cheney et al., 1988)	31	1.7	0.76
Chimpanzee	<i>Pan troglodytes</i>	(Nishida et al., 2003)	26	1.4	0.83
Common buzzard	<i>Buteo buteo</i>	(Kruger and Lindstrom, 2001)	106	3.5	1.14
Subantarctic fur seal	<i>Arctocephalus tropicalis</i>	(Beauplet and Guinet, 2007)	126	2.0	1.17
Cheetah	<i>Acinonyx jubatus</i>	(Kelly et al., 1998)	108	1.7	1.87
Great tit	<i>Parus major</i>	(McCleery et al., 2004)	1795	1.1	1.90
European rabbit	<i>Oryctolagus cuniculus</i>	(von Holst et al., 2002)	197	1.4	2.02
European badger	<i>Meles meles</i>	(Dugdale et al., 2010)	124	1.5	2.18
Northern elephant seal	<i>Mirounga angustirostris</i>	(Le Boeuf and Reiter, 1988)	204	0.8	5.23

OFS, opportunity for selection (standardized variance); N, sample sizes; n.r., not reported. Comparisons are tentative because of differences in assumptions as summarized in (Kelly et al., 1998).

low standardized variance in LRS could be a direct result of the low environmental variability, or it could indicate that the mountain gorillas have low genetic diversity, or it could have arisen by chance (cf. Downhower et al., 1987; Grafen, 1988; Ferguson and Fairbairn, 2001; Shuster and Wade, 2003).

LRS was highly correlated with λ_{ind} as expected because reproduction is distributed fairly evenly throughout adulthood for most females. This mathematical relationship is unlikely to depend upon poaching deaths or other human influences upon the population. The results for λ_{ind} became increasingly insensitive to higher values of LRS, because those higher values mainly arose from greater adult lifespan (Figs. 1 and 2). Similarly, more than 90% of the variance in λ_{ind} among premodern humans arose from differences in LRS, despite diminishing increases in λ_{ind} at higher values of LRS (Kaar and Jokela, 1998). Conversely, the two fitness measures have led to differing conclusions about the optimal age of first reproduction in several species of birds (Brommer et al., 2002). Theoretical studies of lifetime fitness typically focus on λ_{ind} , but empirical studies have often used LRS (Brommer et al., 2004).

Components of LRS

Adult lifespan was the greatest source of variance among the components of LRS, even after adjusting for poaching deaths. In contrast, earlier studies of LRS in mammals and birds had indicated that offspring survival was the most important component, although complete data was not yet available for most long-lived species (Clutton-Brock, 1988). More recent data on large mammals show that longevity has major importance because of its influence on the number of breeding attempts (Berube et al., 1999; Weladji et al., 2006; Petteorelli and Durant, 2007). Among primates, adult lifespan has accounted for 50–70% of the variation in LRS for yellow baboons (Rhine et al., 2000), and 66% of the variation in

LRS for female rhesus macaques (Bercovitch and Berard, 1993).

When we excluded females that died from poaching, offspring survival had relatively little net influence on the variance in LRS, mainly due to the negative covariance with fecundity that arose because females give birth sooner after an infant dies. However, offspring survival could cause more variance in LRS if this population approaches its carrying capacity, even if such variance does not contribute to natural selection. As a population of large mammals approaches its carrying capacity, a decrease in offspring survival typically begins to reduce the growth rate (Eberhardt, 1977; but see Robbins et al., 2009b). Demographic stochasticity can increase the standardized variance in dichotomous variables as their mean decreases (e.g., Fig. 1 of Morris and Doak, 2004), so lower offspring survival could lead to more variance in LRS. Demographic stochasticity can also increase when sample sizes are small (Cabana and Kramer, 1991), so offspring survival may cause more variance in LRS when populations are not growing and/or nonbreeders are rare, because the number of surviving offspring per breeder will be low (Shuster and Wade, 2003; Wade and Shuster, 2004).

Fertility made the smallest contribution to variance in LRS among female mountain gorillas, especially when the analyses were based on interbirth intervals with surviving offspring (IBIS). These results did not include females who died from poaching. Adult lifespans were not significantly correlated with IBIS or the rate of giving birth to offspring that reach weaning age (F_{*S}), so we did not find evidence of a trade-off between reproduction and longevity. Similarly, Ricklefs and Cadena (2007) found no evidence that more frequent reproduction reduced the lifespan of 30 species of captive mammals and birds. Mixed results have emerged from analyses of total births versus postreproductive longevity in humans, perhaps due to lower fertility for unhealthy women, differences in birth spacing, and variability in

the socioeconomic benefits of having large families (Pettay et al., 2005; Hurt et al., 2006; Le Bourg, 2007; Gagnon et al., 2009).

Another potential life history trade-off involves the optimal age for females to begin diverting resources away from their own growth and into reproduction (e.g., Festa-Bianchet et al., 1995; Hill and Hurtado, 1996; Allal et al., 2004; Helle, 2008; Stearns et al., 2010). Females who begin reproducing earlier should typically have more time for subsequent reproduction, unless the early start comes at the expense of harmful effects later in life (Ransome, 1995; Blomquist, 2009b). Studies generally have not discussed how human disturbances would influence this trade-off in other species, but assessments may be complicated by long-term environmental and demographic stochasticity and by differences in female quality (Bercovitch and Berard, 1993; Berube et al., 1999; Neuhaus et al., 2004; Descamps et al., 2006). Larger sample sizes are needed to evaluate this life history trade-off in mountain gorillas.

Dominance rank

Female mountain gorillas with higher dominance rank had significantly longer adult lifespans and greater fitness. Those results excluded females who died from poaching, but the correlations weakened after adjusting for the relationship between rank versus age. In comparison with nepotistic species such as baboons and macaques, the individualistic dominance hierarchies of female mountain gorillas are considered weak, and correlations between rank and different measures of reproductive success have been correspondingly inconsistent (Robbins et al., 2007a,b). Nonetheless, relationships between rank and reproductive success also have been reported for other species with individualistic dominance hierarchies, such as Hanuman langurs (Borries et al., 1991). In nepotistic species and some individualistic species, correlations between rank and reproductive success can mainly reflect social influences such as reproductive suppression and competition for resources (von Holst et al., 2002; Blomquist, 2009a). Those social influences are minimal for female mountain gorillas, as are the other typical sources of environmental variability, so any correlations between their rank and reproductive success are more likely to arise from genetic influences upon differences in physical condition (Robbins et al., 2007b).

Evolution of mountain gorilla life history

The genetic lineage for eastern gorillas diverged from western gorillas approximately 0.9–1.6 million years ago, with gene flow continuing until about 80,000–200,000 years ago (Thalmann et al., 2007). The Virunga mountain gorillas live only 25 km away from the Bwindi mountain gorillas, and 200 km away from eastern lowland gorillas, so it seems likely that gene flow continued among those eastern gorilla populations until barriers to dispersal were created by human agricultural expansion within the past few centuries or millennia (Yamagiwa and Kahekwa, 2001; Robbins et al., 2009b). The adaptive landscape could have shifted repeatedly during the evolution of the modern mountain gorilla, particularly through interglacial oscillations in the climate and the distribution of rain forests (Dupont et al., 2000, 2001; Arnold et al., 2001; Kingsolver and Pfennig, 2007). Overall, however, it seems likely that mountain gorilla life history evolved in an environment

that had fewer humans but more nonhuman predators than this during study, and their habitat encompassed lower elevations that supported a more frugivorous diet (Robbins et al., 2004; Harcourt and Stewart, 2007; Thalmann et al., 2007).

A higher risk of predation has been predicted to favor faster life histories (Promislow and Harvey, 1990; Charnov and Berrigan, 1993; Purvis et al., 2003). Predation by leopards is reported for western lowland gorillas, but the impact on their demography is not well understood (Fay et al., 1995; Robbins et al., 2004). Predation by humans (poaching) reduced the average growth rate of habituated groups in the Virungas from 4.5 to 4.1% during this study, and is probably a major reason why the overall population has grown by only 1% per year (Robbins et al., 2011). Although poaching may have severely reduced the Virunga population prior to this study, the level of poaching during the study could be more comparable to predation pressures throughout the evolutionary history of gorillas.

A more frugivorous diet has been predicted to favor slower life histories, as observed in comparisons among great apes and other primates (Janson and van Schaik, 1993; Wich et al., 2004, 2007). Preliminary evidence suggests a similar correlation among current gorilla populations, notwithstanding any phylogenetic inertia and/or differences in predation pressures (Nowell and Fletcher, 2007; Breuer et al., 2009; Robbins et al., 2009b). It remains unknown, however, whether demographic differences among gorilla populations reflect long term selection pressures on their life history evolution, or if they mainly arise from phenotypic plasticity in response to current ecological conditions including human disturbances (Knott, 2001; Harcourt and Stewart, 2007; Robbins et al., 2009b).

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